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Multifunctional magnetic Fe₃O₄/Cu₂O-Ag nanocomposites with high sensitivity for SERS detection and efficient visible light-driven photocatalytic degradation of polycyclic aromatic hydrocarbons (PAHs)



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HIGHLIGHTS

Fe₃O₄/Cu₂O-Ag nanocomposites (NCs) were designed and synthesized successfully.

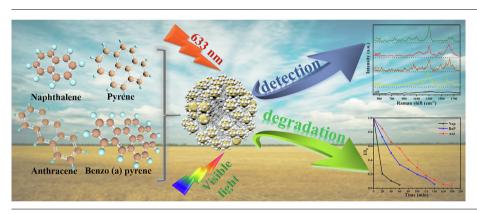
- We realized integration of SERS detection and photocatalytic degradation for PAHs.
- Effect of Ag contents on SERS and catalytic activity of Fe₃O₄/Cu₂O-Ag was studied.
- SERS detection limit of Fe_3O_4/Cu_2O -Ag towards PAHs was as low as 10^{-9} to 10^{-10} M.
- Photocatalytic mechanism of PAHs degradation by Fe₃O₄/Cu₂O-Ag was proposed.

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G R A P H I C A L A B S T R A C T



ABSTRACT

Polycyclic aromatic hydrocarbons (PAHs) with carcinogenic, teratogenic and mutagenic properties are persistent organic pollutants in the environment. Herein, the novel multifunctional Fe₃O₄/Cu₂O-Ag nanocomposites (NCs) have been established for ultra-sensitive surface-enhanced Raman scattering (SERS) detection and visible light-driven photocatalytic degradation of PAHs. Fe₃O₄/Cu₂O-Ag NCs with different amounts of Ag nanocrystals were synthesized, and the effect of Ag contents on SERS performance was studied by finite-difference time-domain (FDTD) algorithm. The synergistic interplay of electromagnetic and chemical enhancement was responsible for excellent SERS sensitivity of Fe₃O₄/Cu₂O-Ag NCs. The limit of detection (LOD) of optimal SERS substrates (FCA-2 NCs) for Nap, BaP, Pyr and Ant was as low as 10⁻⁹, 10⁻⁹ and 10⁻¹⁰ M, respectively. The SERS detection of PAHs in actual soil environment was also studied. Moreover, a simple SERS method was used to monitor the photocatalytic process of PAHs. The recovery and reuse of Fe₃O₄/Cu₂O-Ag NCs were achieved through magnetic field, and the outstanding SERS and photocatalytic performance were still maintained even after eight cycles. This magnetic multifunctional NCs provide a unique idea for the integration of ultra-sensitive SERS detection

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and efficient photocatalytic degradation of PAHs, and thus will have more hopeful prospects in the field of environmental protection.

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1. Introduction

Polycyclic aromatic hydrocarbons (PAHs) are referred to persistent organic pollutants with two or more benzene rings [1-3]. The sources of PAHs include natural sources such as biosynthetic processes of terrestrial and aquatic plants, volcanic ejecta, forest fire and prairie fire, and anthropogenic sources like incomplete combustion of hydrocarbons, wood and fossil fuels, which results in an extremely wide distribution of PAHs in a variety of environments [4,5]. In particular, owing to the high hydrophobic nature of PAHs, the solubility of PAHs is very low in water, which inevitably leads to their strong adsorption in environment, especially in soil, and thus aggravates their accumulation and enrichment in soils [6–8]. However, it is proven that PAHs and their derivatives have teratogenic, mutagenic and carcinogenic risks [9,10]. Even traces of PAHs pose serious threats to human health and cause a variety of diseases, including skin inflammation, kidney and liver damage, metabolic disorders and heart malformations [11,12]. So, it is imperative to design and develop a novel method to realize highly sensitive detection of PAHs in soils and efficient removal of PAHs from soils.

Several advanced technologies have been employed to identify PAHs in soil samples, and the most commonly used methods for sensitive detection of PAHs mainly include gas chromatographymass spectrometry (GC-MS) [13] and high-performance liquid chromatography (HPLC) [14]. Although these conventional methods possess their own advantages, they are often complex, timeconsuming, laborious and costly. Surface-enhanced Raman scattering (SERS) technology, as a simple, fast, cost-effective and efficient technique, is considered to be a promising analytical method for ultra-sensitive analysis of environmental pollutants. Recently, substantial research interests on trace detection of PAHs by SERS are continuously emerging [15–17]. In view of localized surface plasmon resonance (LSPR) characteristics of noble metal plasmonic materials [18], Ag and Au are the most prevalent SERS active substrates for PAHs detection [19,20], given that a considerable electromagnetic field enhancement can be generated at the surfaces of noble metals [21]. But from the perspective of SERS enhancement mechanism, the enhancement of SERS signals depends on two factors - electromagnetic enhancement (EM) as well as chemical enhancement (CE) [22]. Although EM is recognized as a key contribution to SERS enhancement, we should not overlook the role of CE. Combination of semiconductor and noble metal is an effective way to generate CE, which can promote charge transfer (CT) between semiconductor and noble metal and thereby, accelerate CT from SERS substrates to analyte molecules [23]. Moreover, high absorption capacity of SERS substrates to target molecules is another crucial factor to achieve a high SERS enhancement [24]. Generally, PAHs molecules show a quite weak adhesion to metallic surfaces, which is very unfavorable for the SERS detection of PAHs [25,26]. To deal with this issue, much work has focused on the surface functionalization of noble metals with various functional groups [27]. However, these surface functional groups generally produce interfering SERS signals, which will undoubtedly lower the SERS detection sensitivity. Fortunately, apart from the fact that the synergistic effect of EM and CE can enhance the SERS signals of PAHs molecules, the combination of semiconductor and noble metal can facilitate the adsorption of hydrophobic PAHs on SERS substrates [28]. But it is noteworthy that the integration of sensitive detection and efficient removal of PAHs has always been a complicated and challenging research topic so far.

The efficient removal of PAHs from contaminated soils is another challenging task. At present, the effective means for removing PAHs mainly include ozone treatment [29], magnetic solid-phase extraction (MSPE) [30], adsorption [31], bioremediation [32] and photodegradation [33]. Among these various strategies for PAHs removal, photocatalytic degradation technology has attracted extensive concern because of its facile operation, low energy consumption and environmental friendliness. Conventional semiconductor photocatalysts such as ZnO and TiO2 with a wide band gap have been used for the catalytic degradation of PAHs [34,35]. However, these photocatalysts can only be excited by ultraviolet (UV) irradiation, accounting for a small percentage of solar energy (only 4 %), which cannot make full use of solar energy [36]. By comparison, Cu₂O, a kind of narrow band gap semiconductor, can be easily excited by visible light occupying about 43 % of solar energy [37]. Nonetheless, photogenerated electron-hole pairs in Cu₂O tend to recombine, which restricts the photocatalytic efficiency of Cu₂O [38]. Aiming at boosting the separation and transfer of photogenerated electron-hole pairs and hence enhancing the photocatalytic efficiency of Cu₂O, the construction semiconductor-plasmonic metal heterostructures has been proven to be an effective way owing to the appearance of Schottky barrier at the semiconductors/plasmonic metals interfaces [39]. Furthermore, the plasmonic metals can transfer the absorbed light energy to Cu₂O and subsequently prompt Cu₂O to generate more photogenerated electron-hole pairs [40,41]. Therefore, semiconductor/plasmonic metal nanocomposites (NCs) not only can serve as SERS sensors for PAHs, but also can achieve the photocatalytic degradation of PAHs. As far as we know, however, little note has been taken on the integration of ultra-sensitive detection and efficient degradation of PAHs so far.

Recycling of NCs is a prerequisite for sustainable environmental monitoring and remediation. Most frequently used methods for separation of NCs involve centrifugation, filtration and sedimentation, but the procedures of these separation technologies are tedious, time-consuming and render loss of functional properties [42]. Magnetic separation technique possesses advantages of convenient operation and time saving, and superparamagnetic Fe₃O₄ nanocrystals with large specific surface area and high saturation magnetization have been optimal choice for magnetic supports [43]. In order to adhere semiconductor/plasmonic metal to surfaces of Fe₃O₄ nanocrystals, modification of Fe₃O₄ nanocrystals with polymer (e.g. resorcinol-formaldehyde (RF) resins) is indispensable [44]. Uniform RF resin shells are lightly molded on the surfaces of Fe₃O₄ nanocrystals because of the strong interactions between metal ions as well as phenolic hydroxyl groups [45]. In addition, RF resin shells protect Fe₃O₄ nanocrystals from oxidation, aggregation, and shield the interferences of Raman signals originated from Fe₃O₄ nanocrystals.

Inspired by the above ideas, we developed a multifunctional Fe₃O₄/Cu₂O-Ag nanocomposite, which could act as an ultrasensitive SERS sensor and possess effective photocatalyst activity driven by visible light. 4-mercaptobenzoic acid (4-MBA) was chosen as reporter molecule to discuss the effect of Ag contents on SERS activity. The "hot spot" distribution was simulated using finite-difference time-domain (FDTD) algorithm, and corresponding SERS enhancement mechanism was analyzed. Fe₃O₄/Cu₂O-Ag

NCs were served as SERS sensor to realize SERS detection of PAHs in soils. In addition, photocatalytic performance of Fe₃O₄/Cu₂O-Ag NCs for degradation of PAHs under visible light was studied, and photocatalytic mechanism was also proposed. It was proved that the addition of Ag nanocrystals led to formation of Schottky barriers at interfaces between Ag and Cu₂O, which improved the photocatalytic activity of Fe₃O₄/Cu₂O-Ag NCs. Meanwhile, due to the existence of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O-Ag NCs could be conveniently recovered by a magnet and the repeatability of Fe₃O₄/Cu₂O-Ag NCs for SERS detection and photocatalytic degradation of PAHs could be achieved. This work not only enlarges research about SERS and photocatalytic mechanism of multifunctional sensors, but provides a bright future for the tailored design of recyclable and efficient hybrid sensor for detection and degradation of PAHs at the mass production level.

2. Experimental section

2.1. Materials and characterization

The information about materials and characterization are recorded in **Supplementary Materials**.

2.2. Synthesis of Fe_3O_4/Cu_2O and Fe_3O_4/Cu_2O -Ag NCs

As shown in Scheme 1a and 1b, the flower ball-like Fe_3O_4 nanocrystals were prepared by solvothermal method. And Fe_3O_4 @-RF nanocrystals were prepared on the basis of the flower ball-like Fe_3O_4 nanocrystals. The preparation process is listed in Supplementary Materials in detail.

As presented in Scheme 1c, 45 mg of Fe $_3O_4$ @RF NCs was dispersed in deionized water. Next, 5 mL of CuCl $_2$ (0.11 M) was dripped under sonication. Subsequently, 1.8 mL of NaOH (1.0 M) as well as 12 mL of NH $_2$ OH·HCl (0.1 M) were poured and stirred

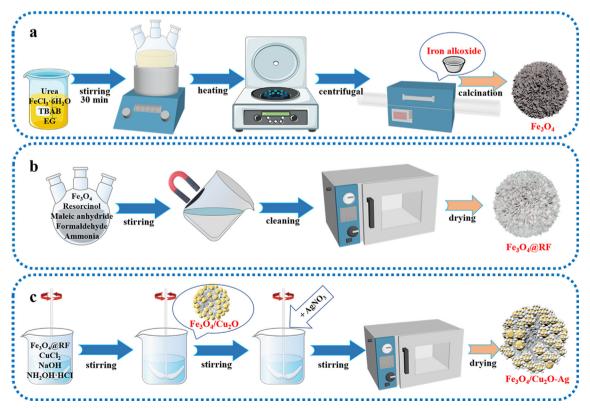
vigorously for 50 min to obtain Fe₃O₄/Cu₂O NCs. Here, NaOH and NH₂OH·HCl were employed as precipitant and reducing agent, respectively. In order to further acquire the Fe₃O₄/Cu₂O-Ag NCs, different volumes of AgNO₃ solution (0.1 M; 1, 2 and 3 mL) were added to the above systems, respectively. After the mixture was stirred for 40 min, the deposit was washed dried. Lastly, the Fe₃O₄/Cu₂O-Ag NCs with different Ag loading amounts were prepared and labeled as FCA-1, FCA-2 and FCA-3, respectively.

2.3. FDTD algorithm method

The information about FDTD software and parameters are described in **Supplementary Materials**.

2.4. SERS measurements

For investigating the influences of Ag content on SERS signals and thus ascertaining an optimal SERS substrate, Fe₃O₄/Cu₂O-Ag NCs with various Ag loading amounts were acted as SERS substrates for detection of reporter molecule (4-MBA). 1 mg of FCA-1, FCA-2 and FCA-3 NCs was added into 3 mL of 4-MBA solution (10⁻⁴ M), respectively. Then, the mixture was blended and shaken on an oscillator without light condition for several hours. The 4-MBA-bound samples were centrifuged, washed with ethanol to dislodge the uncombined 4-MBA molecules and then dried for following SERS detection. SERS spectroscopy was performed at 633 nm laser with an energy density of 10 mW. The SERS spectra of the system were scanned between 800 and 1700 cm⁻¹ and acquisition time was about 10 s. According to SERS results, FCA-2 was determined as optimal SERS substrate, which was chosen as SERS substrate to detect SERS spectra of Nap, BaP, Pyr and Ant. SERS detection process of Nap, BaP, Pyr and Ant was the same as the above SERS detection of 4-MBA. Furthermore, SERS spectra of 4-MBA were acquired at 20 random spots on FCA-2 to evaluate



Scheme 1. Schematic diagram of synthetic steps for Fe₃O₄(a), Fe₃O₄@RF nanocrystals (b), Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs (c).

the uniformity of SERS signal. For obtaining the limit of detection (LOD) of Nap, BaP, Pyr and Ant, these four stock solutions were deliquated to different concentrations (10^{-4} M - 10^{-11} M), respectively. The LOD was measured with the above procedures. To evaluate the practicalities of Fe₃O₄/Cu₂O-Ag NCs, SERS spectra of PAHs in soil were achieved, given that soils were typical carriers containing these persistent organic pollutants.

2.5. Photocurrent response measurements

Standard three-electrode circuit was employed to analyze photocurrent response in a CHI660D electrochemical workstation. Ag/AgCl, platinum wire and Fe $_3$ O $_4$ /Cu $_2$ O-Ag NCs were chosen as reference electrode, counter electrode and working electrode, respectively. Na $_2$ SO $_4$ aqueous solution (0.5 M) was used as electrolyte. 0.003 g of photocatalyst powder and 0.01 g of PVP were blended with 3 μL of absolute ethanol and 3 μL of oleic acid. After continuous ultrasonic blending for 30 min, the slurry was formed and then was evenly coated on the ITO glass. For enhancing the adhesion, the ITO glass was heated to the certain high temperature for several hours.

2.6. Photocatalytic experiments

For investigating the influence of Ag loading on photocatalytic activity and thus determining an optimal photocatalyst, Fe₃O₄/Cu₂O-Ag NCs with different Ag loading amounts were also acted as photocatalysts for catalytic degradation of PAHs. Firstly, 50 mg of Fe₃O₄/Cu₂O, FCA-1, FCA-2 and FCA-3 NCs was put into 100 mL of Nap solution (5 mg/L), respectively. 4 mL of aqueous combination was extracted and leached via a 0.22 mm filter for further analysis in a time interval. In order to prove the fact that Nap was degraded by photocatalysis rather than spontaneous degradation, we monitored degradation process of Nap with or without light. According to photocatalytic results, FCA-2 was identified as an optimal photocatalyst and was used to catalyze the reduction of BaP and Ant. Photocatalytic experiments of BaP and Ant were similar to described photocatalyzed process of Nap.

3. Results and discussion

3.1. Synthesis and characterization of Fe₃O₄/Cu₂O-Ag NCs

Morphologies of iron alkoxide precursors, Fe₃O₄, Fe₃O₄@RF nanocrystals, Fe₃O₄/Cu₂O NCs and Fe₃O₄/Cu₂O-Ag NCs were examined by SEM. As shown in Figure S1a1, iron alkoxide precursors exhibit flower ball-like shapes. Flower ball-like Fe₃O₄ nanocrystals can be obtained by calcining iron alkoxide precursors (Figure \$1a2). Fe₃O₄@RF nanocrystals still maintain flower ball-like structures after Fe₃O₄ nanocrystals are modified with RF layers, and surfaces of the Fe₃O₄ nanocrystals become smoother (Figure S1a3). As presented in Figure S1a4, Cu_2O (\sim 150 nm) truncated octahedron nanocrystals are successfully decorated on Fe₃O₄@RF nanocrystals. Figure S1a5 is SEM image of Fe₃O₄/Cu₂O-Ag NCs. Grain size of Ag nanocrystals is too small (~5 nm), and it is difficult to distinguish them from Fe₃O₄/Cu₂O-Ag NCs. For verifying the existence of Ag nanocrystals, EDS elemental mapping was carried out. According to the EDS elemental analysis results in Figure S1c recorded from the red box in Figure S1a5, elements of Fe, Cu, O and Ag are randomly and homogeneously distributed in Fe₃O₄/Cu₂O-Ag NCs, which strongly confirms the formation of Ag nanocrystals. Apart from SEM images, XRD technology was used for structural characterization of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O NCs and Fe₃O₄/Cu₂O-Ag NCs. As illustrated in **Figure S1b**, typical diffraction peaks of Fe₃O₄ nanocrystals located at 20 values of 30.1°, 35.4°,

 43.0° , 53.4° , 56.9° , and 62.5° are indexed to (220), (311), (400), (422), (511) and (440) crystalline planes of Fe₃O₄ (JCPDS 85-1436), respectively [46]. No other impurities are detected. It is worthwhile to note that it is not sufficient to determine formation of Fe₃O₄ only by XRD, for the reason that Fe₃O₄ and γ -Fe₂O₃ have almost the identical crystal structure [47]. Therefore, Mössbauer spectroscopy measurements were carried out to further confirm existence of Fe₃O₄. Mössbauer spectrum and corresponding Mössbauer spectrum parameters of flower ball-like Fe₃O₄ nanocrystals are presented in Figure S2 and Table S1. Mössbauer spectrum can be fitted into two Zeeman sextets. Sharp and significant lines of magnetic sextets exhibit a double six peak structure, which is the typical characteristic of Fe₃O₄ [48,49]. Hyperfine field is 48.9 and 45.8 Tesla, and the isomer displacement is 0.311 and 0.570 mm/s, corresponding to Fe²⁺ and Fe³⁺ in octahedral interstitial sites and Fe3+ in tetrahedral interstitial sites. After Cu2O nanocrystals are successfully modified on surfaces of Fe₂O₄, four new XRD peaks are founded at 2θ values of 36.4° , 42.2° and 61.2°, which are assigned to (111), (200), and (220) planes of Cu₂O ([CPDS 05-0667) [50]. As for Fe₃O₄/Cu₂O-Ag NCs, additional three XRD peaks are observed at 20 values of 38.0°, 44.2° and 66.5°, which correspond to (111), (200), and (220) planes of Ag (JCPDS 04-0783) [51]. XPS technology was applied to ascertain the valence of elements in Fe₃O₄/Cu₂O-Ag NCs. Full XPS spectra of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs are exhibited in Figure S3. Fe 2p, Cu 2p, Ag 3d, C 1s and O1s are observed and no impurity is found within detection limit of XPS. High-resolution XPS spectra of Fe 2p, Cu 2p and Ag 3d are shown in Figure S4. Fe 2p spectra in Figure S4a display a prominent peak at 710.2 eV and a comparatively weak peak at 723.4 eV for three samples, which are attributed to characteristics of Fe $2p_{3/2}$ and Fe $2p_{1/2}$ in Fe₃O₄ [52,53]. Cu 2p XPS spectra for Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs in Figure S4b show two main peaks at 952.6 and 932.6 eV, which are ascribed to Cu $2p_{1/2}$ and Cu $2p_{3/2}$ of Cu⁺ in Cu₂O [54,55]. As depicted in Figure S4c, peaks of Fe₃O₄/Cu₂O-Ag NCs at 374.2 and 368.2 eV with a spin-orbit splitting of 6 eV are assigned to Ag $3d_{3/2}$ and Ag $3d_{5/2}$ of metallic silver [56]. Notably, binding energies of Cu 2p in Fe₃O₄/Cu₂O-Ag NCs are slightly higher than those in the Fe₃O₄/Cu₂O NCs, which can be attributed to the charge transfer between Cu₂O and Ag [57]. In addition, the modification of Ag on the surfaces of the Fe₃O₄/Cu₂O NCs results in the peak intensity of Cu 2p. Therefore, the results obtained from SEM, XRD, Mössbauer and XPS confirm successful preparation of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O NCs and Fe₃O₄/Cu₂O-Ag NCs.

3.2. Optimization of Fe₃O₄/Cu₂O-Ag NCs substrates

For determining the optimal SERS substrate, the number of Ag on Fe₃O₄/Cu₂O NCs was adjusted by adding various amounts of AgNO₃. As exhibited in Fig. 1a1 and a2, when the addition amount of AgNO₃ changes from 1 to 2 mL, the surfaces of Cu₂O become rougher owing to the increase of Ag absorbed on the surfaces. Self-aggregation of Ag can be observed obviously when the amount of AgNO₃ further increases to 3 mL (Fig. 1a3) [58]. To reveal changes of microstructure and elemental distributions when increasing Ag contents, Fe₃O₄/Cu₂O-Ag NCs were measured by TEM, high angle annular dark-field (HAADF) STEM and EDS elemental mapping. It can be seen from Fig. 1b and 1c that Fe₃O₄ nanocrystals with a diameter of about 1.5 µm are located at the center of Fe₃O₄/Cu₂O NCs and are surrounded with Cu₂O and Ag nanocrystals. Ag nanocrystals on surfaces of Fe₃O₄/Cu₂O NCs increase significantly with increase of Ag addition. However, excessive Ag loading results in aggregation and coalescence of Ag nanocrystals. Given that SERS enhancement ability is dependent upon the number of noble metal nanocrystals and the spacing between noble metal nanocrystals [59], SERS performance of

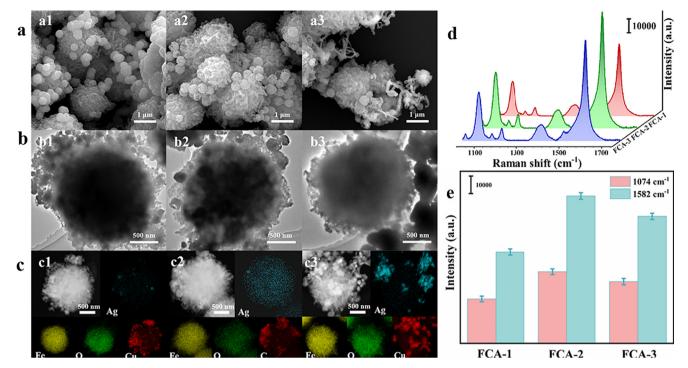


Fig. 1. SEM, TEM, HAADF-STEM images and corresponding EDS elemental mapping results (Ag, Fe, O and Cu) of Fe₃O₄/Cu₂O-Ag NCs (FCA-1) (a1, b1, c1), (FCA-2) (a2, b2, c2) and (FCA-3) (a3, b3, c3); SERS spectra of 4-MBA (10⁻⁴ M) on FCA-1, FCA-2 and FCA-3 NCs (d) and the histograms of SERS signals of 4-MBA at 1074 cm⁻¹ and 1582 cm⁻¹ (e).

Fe₃O₄/Cu₂O-Ag NCs with different Ag contents was evaluated. 4-MBA (10⁻⁴ M) was acted as a reporter molecule to explore relationship between Ag content and SERS signal intensity, as presented in Fig. 1d. Detailed band assignments of 4-MBA SERS spectra are presented in **Table S2** [60]. Fig. 1e illustrates SERS intensity of 4-MBA adsorbed on the Fe₃O₄/Cu₂O-Ag NCs at 1074 and 1582 cm⁻¹. And error bar represents standard deviation obtained from five independent tests. The SERS intensity of 4-MBA rises up and then declines with increasing content of Ag nanocrystals (Fig. 1e), and 4-MBA adsorbed on FCA-2 NCs has a higher SERS intensity compared with FCA-1 and FCA-3 NCs. This phenomenon is to be expected and will be discussed below.

3.3. Mechanism of SERS enhancement by Fe₃O₄/Cu₂O-Ag NCs

EM mechanism is generally considered to play a major part in SERS enhancement, which mainly derives from excitation of LSPR on noble metal nanostructures [61–63]. Especially, owing to the coupling effect between noble metal nanocrystals, stronger electromagnetic fields will be generated in/between adjacent noble

metal nanocrystals, called "hot spots", so as to considerably enhance SERS signals of target molecules [64]. One clear reason is that multiple collisions between the photons excited by incident light and noble metals nanocrystals can improve SERS signal intensity [65]. As displayed in Fig. 2a, the electromagnetic field intensity dispersion of FCA-2 NCs under periodic boundary conditions was simulated and analyzed by FDTD method, where a large number of hot spots are located in the gap between Ag nanocrystals. Because number of Ag nanocrystals on surface of FCA-2 NCs is more than that of FCA-1 NCs, FCA-2 NCs have more "hot spots". Given that SERS signal intensity relies on amount of "hot spots" in SERS substrates, it is reasonable that FCA-2 NCs have better SERS performance than FCA-1 NCs [66.67]. Furthermore, unexpected decrease of SERS signal on FCA-3 NCs can be attributed to the excessive addition of Ag nanocrystals, which results in agglomeration of Ag nanocrystals and thus decreases the number of "hot spots" [59]. CE also makes contribution to SERS enhancement through CT between adsorbed molecules and SERS substrates [68]. It is well known that SERS peaks of 4-MBA mainly include a₁ symmetric vibration modes and b₂ asymmetric vibration modes,

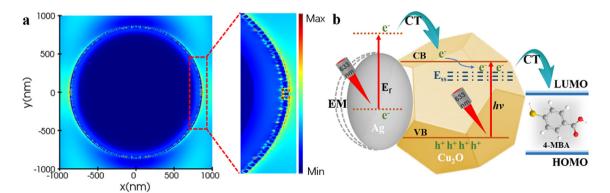


Fig. 2. Electric field distribution of FCA-2 NCs simulated by FDTD method (a) and schematic diagram of CT between 4-MBA and Fe₃O₄/Cu₂O-Ag NCs (b).

in which enhancement of b2 vibration modes originates from CT between substrates and 4-MBA [69,70]. The b₂ vibration bands of 4-MBA absorbed on Fe₃O₄/Cu₂O-Ag NCs at 1144 cm⁻¹ are clearly observed in Fig. 1d. In order to account for CT process between Fe₃O₄/Cu₂O-Ag NCs and 4-MBA, the CT mechanism was proposed, as shown in Fig. 2b. The deposited Ag nanocrystals on Cu₂O surface can be excited by LSPR adsorption. Photoexcited electrons can be injected into the conduction band (CB) of Cu₂O and/or then be vibratingly relaxed to the surface energy levels (E_{SS}) of Cu₂O, and eventually be diverted to the LUMO of 4-MBA. For the present Ag-Cu₂O-molecular system, the CB of Cu₂O acts as a "bridge" between Ag (donor) and 4-MBA (receptor), achieving a "donor bridge receptor" CT mode [71,72]. A small quantity of Ag nanocrystals cannot provide enough photoexcited electrons. On the contrary, excess Ag nanocrystals will cover the surfaces of Cu₂O and lead to insufficient light absorption. Therefore, we conclude that the appropriate amount of Ag nanocrystals is crucial to achieve better SERS performance based on both EM and CE mechanism. Herein, an extremely high enhancement factor (EF) of 3.08×10^6 was achieved according to the following equation [73,74]:

$$EF = \frac{I_{SERS}N_{Solid}}{I_{Solid}N_{SERS}} = \frac{I_{SERS}}{I_{Solid}} \times \frac{\rho \times h \times S_{laser} \times N_A/M}{S_{laser}/S_{4-MBA}}$$

Specific EF calculation process and Raman intensity (Figure S5) of solid 4-MBA powders are recorded in Supplementary Materials.

3.4. Stability and reproducibility of Fe₃O₄/Cu₂O-Ag NCs

Stability of substrates and reproducibility of SERS signals of target analytes are vitally important in actual application. So as to assess stability of SERS substrates, FCA-2 NCs were stored in a centrifuge tube for 0, 10, 20 and 30 days. Histogram of average SERS intensity of the 4-MBA at 1074 and 1582 cm⁻¹ on FCA-2 NCs stored for different times is shown in Figure S6a. There is no significant reduction in SERS intensity of 4-MBA at 1074 and 1582 cm⁻¹ even after a month, indicating Fe₃O₄/Cu₂O-Ag SERS substrates have good storage stability. For assessing reproducibility of SERS signals, SERS spectra of 20 random points on FCA-2 NCs functionalized with 10⁻⁴ M 4-MBA were obtained (Figure S6b). Corresponding relative standard deviation (RSD) of 4-MBA at 1074 as well as 1582 cm⁻¹ was calculated by the linear regression method (Figure S6c). Results suggest that all the SERS signals of 4-MBA are highly consistent, and RSD values of 4-MBA at 1074 and 1582 cm⁻¹ are 9.73 and 10.82 %, which implies that Fe₃O₄/Cu₂O-Ag SERS substrates have high uniformity and excellent spectral reproducibility in a largescale area.

3.5. SERS applications of Fe₃O₄/Cu₂O-Ag NCs

Four kinds of PAHs (Nap, BaP, Pyr and Ant; $10^{-11} \sim 10^{-4}$ M) were selected as probe molecules to assess applicability of FCA-2 NCs as SERS substrates for quantitative detection of PAHs. Fig. 3 shows that the characteristic peaks of Nap, BaP, Pyr and Ant can be observed clearly, and LOD of Nap, BaP, Pyr and Ant is as low as 10^{-9} , 10^{-9} , 10^{-9} and 10^{-10} M, respectively. In addition, relationship between the concentrations of Nap, BaP, Pyr and Ant adsorbed on FCA-2 NCs and the corresponding SERS intensity at 1379, 1555, 1599 and 1380 cm $^{-1}$ are illustrated in Figure S7. SERS intensity for four kinds of PAHs follows linear relationship versus the logarithm of the concentrations of PAHs and correlation coefficients (R^2) are more than 0.97, which indicates that Fe₃O₄/Cu₂O-Ag NCs are ultra-sensitive SERS substrates and can realize quantification of different kinds of PAHs down to 10^{-9} M.

In order to verify application effect of Fe₃O₄/Cu₂O-Ag NCs for PAHs detection in actual environment, soil samples taken from Jilin

Normal University (Changchun China) were detected. The soil samples were heated at 500 °C for 3 h to avoid interferences of other soil components on SERS spectra. As exhibited in Fig. 4a, heattreated soil (20 mg) was added to a centrifuge tube and mixed with four kinds of PAHs (Nap, BaP, Pyr and Ant 10^{-4} M, 3 mL) for 2 h, respectively. Subsequently, 1 mg of FCA-2 NCs was added to the above mixtures, respectively. After shaking with an oscillator in dark overnight, PAHs labeled FCA-2 NCs were separated and collected from the mixed solution by a magnet. Finally, PAHs labeled FCA-2 NCs were transferred to clean slides, and SERS signals of the four samples were recorded excited with 632.8 nm laser. As shown in Fig. 4b, no obvious SERS signals are obtained for the heat-treated soil samples. In contrast, the SERS signals of four kinds of PAHs can be clearly detected, demonstrating that Fe_3O_4/Cu_2O -Ag NCs are suitable for the SERS detection of PAHs in the actual environment.

3.6. Photocatalytic performance of Fe₃O₄/Cu₂O-Ag NCs

Efficient photocatalytic degradation of PAHs is another key development goal while achieving ultra-sensitive SERS detection. Given that the separation and transfer ability of photogenerated carriers largely determine photocatalytic performance of photocatalysts, transient photocurrent measurements were implemented to estimate excitation and transfer of photoexcited electron-hole pairs in Fe₃O₄/Cu₂O-Ag NCs [75]. Transient photocurrent responses of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs with different Ag contents (FCA-1, FCA-2 and FCA-3) are presented in Fig. 5a. Under visible light illumination, all samples have quick and reproducible photocurrent response, and photocurrent density decreases rapidly to zero after cessation of light excitation. In addition, Fe₃O₄/Cu₂O-Ag NCs have higher photocurrent response than Fe₃O₄ nanocrystals and Fe₃O₄/Cu₂O NCs, which indicates that introduction of Ag can enhance the separation and transfer ability of electron-hole pairs [57]. It is a remarkable fact that the photocurrent density of Fe₃O₄/Cu₂O-Ag NCs increases firstly and then decreases with increase of Ag contents, and reaches a maximum for FCA-2 NCs. A reasonable explanation is that Ag nanocrystals can enhance light absorption owing to LSPR effect of Ag nanocrystals [76,77], but excess Ag nanocrystals are likely to cover surfaces of Cu₂O entirely, which will inhibit generation of photoexcited carriers and thus lead to poor photocatalytic activity [78]. Photoluminescence (PL) spectroscopy was used to further assess recombination of photoexcited carriers for Fe₃O₄/Cu₂O-Ag NCs. As depicted in Fig. 5b, FCA-2 NCs exhibit the lowest PL intensity owing to slowest recombination of charge carries, and this trend is in accordance with above transient photocurrent response result, which provides further evidence that FCA-2 NCs have the highest carrier separation and transfer efficiency [79].

UV–vis DRS was used to explore the effect of Ag nanocrystals on photocatalytic performance of Fe_3O_4/Cu_2O NCs. As illustrated in Fig. 5c, absorption peaks (450–500 nm) of both Fe_3O_4/Cu_2O and FCA-2 NCs appear in visible light region. With the addition of Ag nanocrystals, FCA-2 NCs show stronger visible light absorption compared with Fe_3O_4/Cu_2O NCs, which is ascribed to SPR effect of Ag nanocrystals [80]. Values of band gap energy (E_g) of Fe_3O_4/Cu_2O and FCA-2 NCs were calculated according to experiential Tauc's formula[81]:

$$(\alpha h v)^2 = A(h v - -E_g)$$

Where E_g , α , hv and A stand for band gap energy, absorbance, photon energy and constant, respectively. As exhibited in Fig. 5d, the E_g is determined from intercept of tangent of $(\alpha hv)^2$ versus hv curves. Accordingly, E_g of Fe₃O₄/Cu₂O and FCA-2 NCs is computed to be around 2.32 and 2.23 eV, respectively. Above results

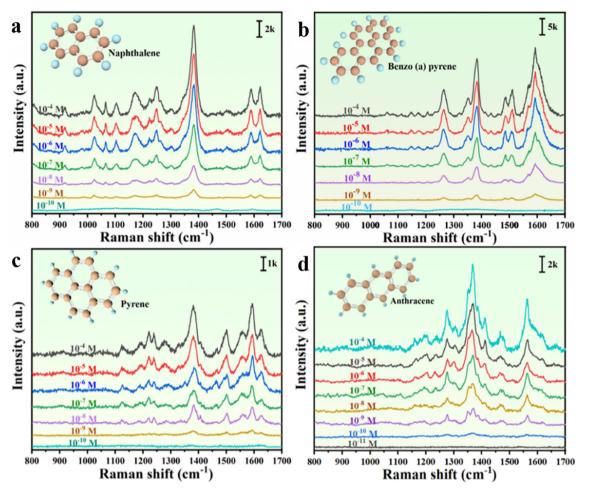


Fig. 3. SERS spectra of Nap (a), BaP (b), Pyr (c) and Ant (d) with different concentrations ($10^{-11} \sim 10^{-4}$ M) adsorbed on FCA-2 NCs.

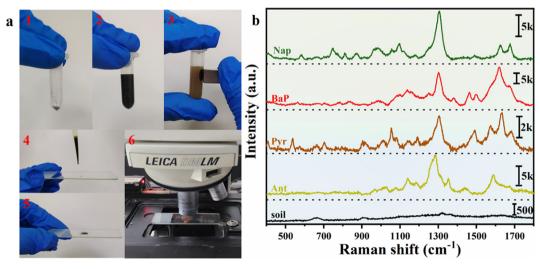


Fig. 4. Operation procedure of SERS detection of PAHs in soil using FCA-2 NCs as SERS substrates (a) and SERS spectra of Nap, BaP, Pyr, Ant and heat-treated soil (b).

indicate that the Fe_3O_4/Cu_2O -Ag NCs (FCA-2 NCs) with a reduced bandgap of 0.09 eV result in an increased visible light response [82].

For intuitively reflecting the catalytic performance of Fe₃O₄/Cu₂O with various loading amounts of Ag nanocrystals, photocatalytic degradation of Nap solution (5 mg/L) catalyzed by Fe₃O₄/Cu₂O, FCA-1, FCA-2 and FCA-3 NCs was monitored by SERS spectra

(Fig. 5e). I_0 is regarded as initial SERS intensity of Nap solution at 1379 cm⁻¹ and I is SERS intensity after irradiation under visible light, respectively. It can be observed that all I/I_0 values decrease with the rise of illumination time and the I/I_0 value of Fe₃O₄/Cu₂O-Ag NCs decline faster than that of Fe₃O₄/Cu₂O NCs, suggesting addition of Ag nanocrystals promotes degradation of Nap. Moreover, FCA-2 NCs exhibit the highest photocatalytic activity

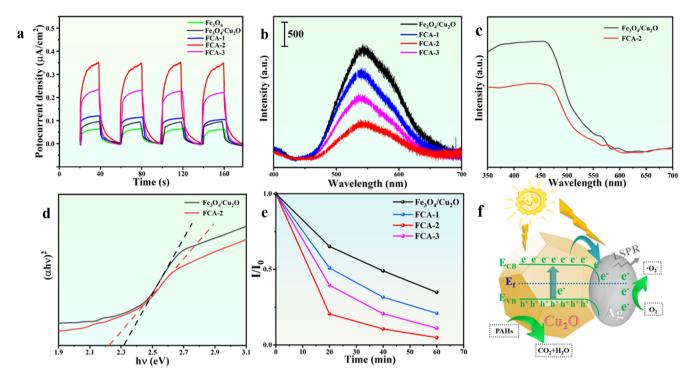


Fig. 5. Photocurrent response of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O, FCA-1, FCA-2 and FCA-3 NCs (a); PL spectra of Fe₃O₄/Cu₂O, FCA-1, FCA-2 and FCA-3 NCs (b); UV-vis DRS (c) and relative band gap energy (d) of Fe₃O₄/Cu₂O and FCA-2 NCs; I/I_0 of SERS signals at 1379 cm⁻¹ of Nap (5 mg/L) versus reaction time in the presence of Fe₃O₄/Cu₂O, FCA-1, FCA-2 and FCA-3 NCs (e) and schematic of photocatalytic degradation mechanism of Fe₃O₄/Cu₂O-Ag NCs for PAHs under visible light illumination (f).

as expected, and degradation rate of Nap by FCA-2 NCs exceeds 95 % after 60 min of illumination. In order to rule out the possibility of spontaneous degradation of Nap, we monitored the degradation process of Nap with or without light, and the histogram of SERS intensity of Nap at 1379 cm⁻¹ is shown in **Figure. S8**. Catalytic degradation of Nap is achieved only with light irradiation, and slight decrease of SERS intensity with no-light condition is attributed to trace adsorption of nanocatalysts.

3.7. Mechanism of photocatalytic degradation by Fe₃O₄/Cu₂O-Ag NCs

On the basis of points discussed above, a possible photocatalytic degradation mechanism of Fe₃O₄/Cu₂O-Ag NCs for PAHs was proposed in Fig. 5f. Once two kinds of nanomaterials (Cu₂O and Ag) with different work functions come into contact with each other, a Schottky barrier forms at Cu₂O-Ag interface [83,84]. Electrons will migrate from Ag with lower work function to Cu₂O with higher work function until Fermi level equilibration takes place. Meanwhile, shift-down of Fermi level of Ag and the shift-up of Fermi level of Cu₂O will result in energy band bending downward. When Fe₃O₄/Cu₂O-Ag NCs are excited by visible light, the photoexcited electrons will transfer from VB of Cu₂O to CB, while leaving an equal number of holes in VB to form the photoexcited electronhole pairs. Considering the CB of Cu₂O is higher than that of new equilibrium Fermi level, photoexcited electrons will continue to transfer from CB of Cu₂O to Ag. However, photoexcited electrons in Ag nanocrystals cannot be migrated to Cu₂O because of the blocking effect of Schottky barrier [85]. Therefore, Ag nanocrystals serve as electron traps to efficiently inhibit the recombination of electron-hole pairs. Photoexcited electrons of Ag surfaces can be captured by the adsorbed O₂ to generate superoxide anion radicals $(\cdot O_2^-)$ with strong oxidative capacity for degradation of PAHs [78]. In addition, the holes left in VB of Cu₂O lead to direct degradation of PAHs molecules. It is worth noting that Ag nanocrystals with LSPR properties can enhance carrier concentration by facilitating

generation of photoexcited electrons and holes, which is also beneficial for degrading the PAHs [86]. In summary, Ag nanocrystals not only promote the formation of active radicals, but also improve light utilization in Fe₃O₄/Cu₂O-Ag NCs. Therefore, it is reasonable that Fe₃O₄/Cu₂O-Ag NCs present better photocatalytic performance than Fe₃O₄/Cu₂O NCs.

3.8. Photocatalytic applications of Fe₃O₄/Cu₂O-Ag NCs

So as to evaluate applicability of Fe₃O₄/Cu₂O-Ag NCs for visible light-driven degradation of different PAHs, in addition to Nap, BaP and Ant were also selected as degradation targets to perform same photocatalytic experiments under visible light irradiation. SERS spectral changes at various times during photocatalytic process were monitored by Raman spectrometer. As presented in Figure S9, it is detected that intensity of SERS peaks of Nap, BaP and Ant decreases in different degrees with increasing irradiation time. Especially, it can also be seen that the position and shape of SERS peaks of PAHs change slightly in the photocatalytic process, which may be due to formation of complex intermediates in the photocatalytic process [87,88]. In order to intuitively reflect changes in SERS signal intensity, the temporal variations of I/I_0 for SERS peaks of Nap (1379 cm⁻¹), BaP (1555 cm⁻¹) and Ant (1380 cm⁻¹) catalyzed by FCA-2 NCs are plotted in Fig. 6a. Characteristic SERS signals of Nap, BaP and Ant almost completely disappear within 60, 140 and 180 min, respectively. Furthermore, the degradation efficiency of Nap, BaP and Ant was calculated by plotting natural logarithm of I/I_0 versus irradiation time, as shown in Fig. 6b. About 80 % of Nap can be degraded on FCA-2 NCs within 20 min, while degradation of same amount of BaP and Nap is achieved within 100 and 120 min, respectively. Although FCA-2 NCs exhibit different catalytic activities towards three PAHs, real-time monitoring of degradation of PAHs catalyzed by FCA-2 NCs clearly indicates that Fe₃O₄/Cu₂O-Ag NCs can achieve catalytic degradation of multiple PAHs.

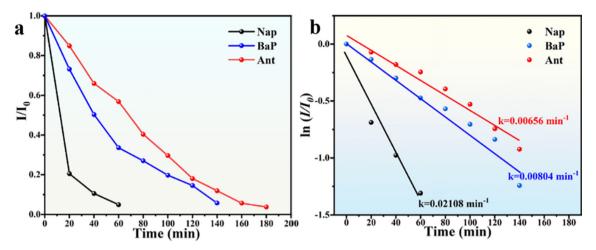


Fig. 6. The plots of l/l_0 for the SERS peaks of Nap (1379 cm⁻¹), BaP (1555 cm⁻¹) and Ant (1380 cm⁻¹) catalyzed by FCA-2 NCs versus illumination time (a) and the corresponding plots of $ln(l/l_0)$ versus reaction time (b).

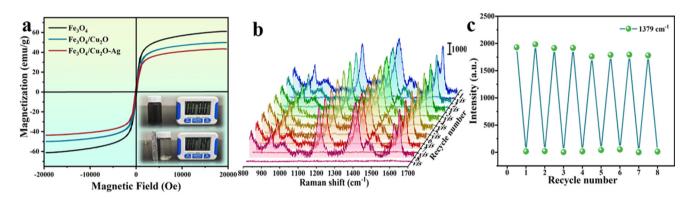


Fig. 7. Magnetic hysteresis (M-H) loops of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs (The inset is photograph of Fe₃O₄/Cu₂O-Ag NCs scattered in water before and after magnetic separation) (a); SERS spectra of Nap (5 mg/L) (b) and their SERS intensities at 1379 cm⁻¹ (c) during eight cycles of visible light-driven photocatalysis.

3.9. Magnetic applications of Fe₃O₄/Cu₂O-Ag NCs

Magnetic properties of Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs (FCA-2 NCs) were tested at room temperature, and their magnetic hysteresis (M-H) loops are verified in Fig. 7a. It is found that Fe₃O₄ nanocrystals, Fe₃O₄/Cu₂O and Fe₃O₄/Cu₂O-Ag NCs possess superparamagnetic, and their saturation magnetization (Ms) values are 61.0, 50.8 and 41.2 emu·g⁻¹, respectively. Compared with Fe₃O₄ nanocrystals, decreased Ms value of Fe₃O₄/ Cu₂O and Fe₃O₄/Cu₂O-Ag NCs is attributed to increased mass of Cu₂O nanocrystals and Cu₂O-Ag NCs [89]. However, even though Fe₃O₄/Cu₂O-Ag NCs have lowest Ms value, they are collected by the magnet within 14 s, as shown in the inset of Fig. 7a. Excellent magnetic response property endows Fe₃O₄/Cu₂O-Ag NCs with capability of recovery and reuse. So as to further assess the stability and reusability of Fe₃O₄/Cu₂O-Ag NCs, sequential SERS detection and photodegradation experiments of Nap (5 mg/L) adsorbed on Fe₃O₄/Cu₂O-Ag NCs were carried out. As reflected in Fig. 7b, the SERS peaks of Nap disappear after visible light irradiation, and Fe₃O₄/Cu₂O-Ag NCs still maintain an excellent SERS activity even after 8 cycles. To intuitively show change of SERS intensity, the SERS intensities of Nap at 1379 cm⁻¹ after eight cycles for detection/photodegradation of Nap are exhibited in Fig. 7c. The Fe₃O₄/ Cu₂O-Ag NCs almost maintain their initial SERS activity in the first 4 cycles. Slight decrease of the SERS signals after 4 cycles may be responsible for the loss of Fe₃O₄/Cu₂O-Ag NCs or reduction of their adsorption capacity during cycling process. However, they still maintain nearly 90 % of the initial SERS activity even after 8 cycles. To sum up, it can draw a conclusion that our proposed Fe_3O_4/Cu_2O_8 NCs are fully recycled and thus have great potential for applications in efficient detection and catalytic degradation of PAHs.

4. Conclusions

In conclusion, magnetic multifunctional Fe₃O₄/Cu₂O-Ag NCs for ultra-sensitive SERS detection and efficient photocatalytic degradation of PAHs have been successfully developed. Ag amounts on surfaces of Fe₃O₄/Cu₂O-Ag NCs were adjusted and 4-MBA was chosen as reporter molecule to investigate the effect of Ag contents on SERS signal intensity. It was observed that SERS intensity of 4-MBA first increased and then decreased when increasing content of Ag nanocrystals, which depended on the numbers of "hot spots" as well as photoexcited electrons in SERS substrates. In addition, little changes in SERS intensity of 4-MBA were found even after a month, suggesting that the Fe₃O₄/Cu₂O-Ag substrates had good stability. Four kinds of PAHs (Nap, BaP, Pyr and Ant) were used to evaluate applicability of Fe₃O₄/Cu₂O-Ag as SERS substrates. When the concentrations of PAHs ranged from 10⁻¹¹ to 10⁻⁴ M, SERS intensity of Nap, BaP, Pyr and Ant followed linear relationship versus the logarithm of the concentrations of PAHs, and LOD of Nap, BaP, Pyr and Ant was as low as 10^{-9} , 10^{-9} , 10^{-9} and 10^{-10} M, respectively. Sensitive SERS detection of PAHs in actual soil environment was also proved. Furthermore, photocatalytic performance of Fe₃O₄/

Cu₂O-Ag NCs for degradation of PAHs was evaluated. Compared with Fe₃O₄/Cu₂O NCs, Fe₃O₄/Cu₂O-Ag NCs exhibited higher photocatalytic activity as a consequence of the formation of Schottky barrier at the Cu₂O-Ag interface. Under visible light irradiation, Nap, BaP and Ant were almost completely degraded within 60, 140 and 180 min, respectively. Fe₃O₄/Cu₂O-Ag NCs possessed superparamagnetic properties, and the Ms value was 41.2 emu·g⁻¹, which endowed them with the ability to be recycled. This research not only provides a novel avenue for the SERS detection and visible light-driven photocatalytic degradation of PAHs, but also shows great potential in ensuring environmental safety to protect public health.

CRediT authorship contribution statement

Jie Huang: Investigation, Methodology, Formal analysis, Writing – original draft. **Tianxiang Zhou:** Investigation. **Wenshi Zhao:** Formal analysis. **Sicheng Cui:** Conceptualization. **Rui Guo:** Data curation. **Dan Li:** Visualization. **Naveen Reddy Kadasala:** Validation. **Donglai Han:** Methodology. **Yuhong Jiang:** Funding acquisition. **Yang Liu:** Conceptualization, Funding acquisition, Writing – review & editing, Supervision. **Huilian Liu:** Supervision.

Data availability

Data will be made available on request.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary material

Supplementary data to this article can be found online at https://doi.org/10.1016/j.jcis.2022.08.036.

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